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1 Engaging urban stakeholders in the sustainable management of arthropod pests

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20 [Abstract](#)

21 The management of arthropods in urban environments is complex. Although there are species that
22 threaten human health and property, there are also extensive communities of beneficial species that
23 need to be conserved. Current management of arthropod pests in cities relies heavily on the use of
24 synthetic chemicals, which have a range of potential environmental and health impacts. In order to
25 mitigate the impacts of insecticides, urban stakeholders need to be encouraged to reduce reliance on
26 chemical control and adopt more ecologically sustainable approaches. Integrated Pest Management
27 (IPM) has been globally successful in managing pests in agriculture but has yet to be broadly
28 practiced in urban systems. Here, we address the global problem of lack of IPM uptake in urban areas.
29 We summarise current arthropod management practices, with comparisons drawn between the
30 management of pests in urban and agricultural systems, and highlight the benefits of IPM. We then
31 give examples of successful IPM to demonstrate the useful implementation strategies and identify key
32 barriers to the adoption of this approach in urban systems. In particular, the high diversity of
33 stakeholder interests and management practices is a key barrier to overcome in cities, along with lack
34 of awareness of the benefits and implementation strategies of IPM, little emphasis on monitoring
35 pests, restrictions in time/resources; and social factors such as negative public perceptions of insects
36 and policy regulations. We offer suggestions for overcoming these barriers in the hope of encouraging
37 greater application of sustainable arthropod pest management practices for all urban stakeholders.

38

39 [Keywords](#)

40 Integrated pest management, urbanisation, arthropod, management, community

41 [Key Message](#)

- 42 • Despite extensive benefits, integrated pest management (IPM) has had little uptake in urban areas.
- 43 • Barriers include limited integration between disciplines and regulatory bodies, a lack of knowledge
44 of urban pest ecology, and diverse stakeholder interests.
- 45 • Overcoming these barriers requires: comprehensive monitoring systems; IPM compatible tools;
46 support for commercial operators to be agents for change; and the development of support
47 networks for all stakeholders.
- 48 • Adoption of IPM in cities will reduce environmental and public health impacts, redefine how
49 residents relate to arthropods and encourage sustainable management of urban ecosystems.

50 [Author Contribution Statement](#)

51 E.L. conceived the idea and assembled all co-authors. E.L., T.L., C.W., M.W., and M.S. all
52 contributed equally to formation of the outline and writing the content of this review.

53

54 1. Introduction

55 Building cities drastically modifies natural habitats, which adversely impacts arthropod communities
56 and ultimately drives the localised extinction of many species (McKinney 2008). Some arthropods are
57 able to adapt to these changes, leading to increases in abundance and, in some cases, the establishment
58 of the species as a pest. A pest is an anthropocentric term, and simply means a disruptive plant or
59 animal that causes annoyance or damage in a given context. In an agricultural system this primarily
60 refers to organisms that damage crops, affect livestock health or stored product quality. In an urban
61 setting, the main pests of concern are usually those that cause nuisance, present a human health risk
62 through the transmission of pathogens, inflict structural damage and impact local plants (Robinson
63 1996). While the disservices associated with arthropod populations may require active management,
64 strategic responses should also consider the ecosystem services provided by arthropods in the urban
65 environment. There is also often a mismatch between the public perception of arthropods and the pest
66 impacts they pose. This results in the use of insecticides to control arthropods that do not necessarily
67 pose a substantial economic or public health threat (Shrewsbury and Leather 2012). For the purpose
68 of our review, pests are organisms that have a direct economic impact on public health, urban
69 vegetation or built structures, rather than arthropods that may simply be perceived by some people as
70 a pest but cause no direct damage (New 2015).

71 Any arthropod species can become a pest in certain contexts, even perceived beneficial species,
72 depending on their ecological characteristics. Arthropod pests typically arise when humans increase
73 their proximity to pest species' habitats, where modification of the environment increases the numbers
74 of a native species, or where a non-native species is introduced (New 2015). In cities, some
75 arthropods adapt to, and thrive in, the newly created habitats (e.g. urban waterways, dwellings, waste)
76 and once established, can cause damage to: structures, stored products, food production and both
77 native and cultivated vegetation; as well as presenting significant health risks. The population
78 dynamics of urban arthropod pests and their predators are influenced by a range of factors including
79 land use (Burkman and Gardiner 2014, Nelson and Forbes 2014), climate (Meineke, Dunn et al. 2014)
80 and biological community structure such as increases in prey or loss of predators (Faeth, Warren et al.

81 2005). Globalisation also increases the opportunities for exotic pest species to be introduced to new
82 urban areas (Benedict, Levine et al. 2007).

83 Maintaining the function of global urban ecosystems is essential. To achieve effective management of
84 these systems, urban stakeholders should adopt a preventative approach based on long-term
85 investment and evidence-based decisions that reduce pest damage and public health risks while
86 conserving the diversity of beneficial arthropods (New 2018). This will be a critical challenge with
87 the trend of “greening cities” and the subsequent management of urban green spaces (Aronson,
88 Lepczyk et al. 2017) and consideration will need to be given to how the beneficial and pest arthropods
89 associated with these habitats are managed.

90 In order to encourage local authorities and the community to reduce their reliance on chemical control
91 and increase uptake of sustainable arthropod pest management practices in urban areas, our study
92 aims to a) give an overview of current management, b) explain Integrated Pest Management (IPM)
93 and how it could be adapted for use in urban pest management practice c) provide examples of
94 successful IPM strategies and outline strategies that could be used in urban areas, and d) describe the
95 key barriers to uptake and changes in practice needed in order to implement IPM strategies in cities.
96 This paper also identifies knowledge gaps and explores the need for partnerships between local
97 communities, government, industry and research in order to address these problems.

98 2. Management of arthropods in cities

99 Urban pest control strategies are traditionally reactive and often locally-applied, rather than focused
100 on long-term integrated management of pest arthropod populations (Knipling 1998; Varga 2017). This
101 strategy is often prompted by the need for short-term control of arthropods posing a public health risk
102 or immediate threat to property or vegetation. The diversity of habitat types and municipal systems in
103 urban areas, as well as different resource requirements and life history of urban arthropod pests,
104 makes broad scale management a challenge.

105 Since the 1940s the use of chemicals to control pests has increased in urban areas because it is
106 considered cheap, easily accessible and effective (Dahlsten and Hall 1999; Grube et al. 2011).

107 Household and municipal pest control is still dominated by high levels of pesticide use and while the

108 reduction in pest arthropods of public health importance (e.g. mosquitoes) is of benefit, consideration
109 should also be given to less positive outcomes of insecticide use in both the environment and
110 community (Julien, Adamkiewicz et al. 2008). These include significant environmental impacts such
111 as runoff into waterways (Amweg, Weston et al. 2006, Holmes, Anderson et al. 2008, Weston, Asbell
112 et al. 2011, Wittmer, Scheidegger et al. 2011) and human health impacts ranging from developmental
113 disorders to chronic health problems (Ma, Buffler et al. 2002, Rauh, Perera et al. 2012, Hernández,
114 Parrón et al. 2013, Narayan, Liew et al. 2013). However, a balance is often required where the risk to
115 human health of insecticides is exceeded by the risks posed by arthropods of medical importance (e.g.
116 vectors of pathogens) (Peterson, Macedo et al. 2006). While the use of insecticides is regulated by
117 government authorities in many parts of the world (Li and Jennings 2017), even where human health
118 impacts may be low, there remains concern regarding the non-target impacts of broad-scale
119 insecticide use and subsequent ecological impacts. These concerns, together with challenges posed by
120 increasing rates of insecticide resistance in key pest species of public health concern (Deming et al.
121 2016), have driven research into alternative strategies of mosquito management. The release of
122 laboratory reared mosquitoes either carrying pathogens that inhibit pathogen transmission (Hoffmann
123 et al. 2011) or reproductive success (Rasic et al. 2014), or are genetically modified to drive reductions
124 in local mosquito abundance (Carvalho et al. 2015) are strategies gaining increasing interest from
125 local authorities. These approaches may substantially reduce the non-target impacts of broad spectrum
126 insecticides but require substantial investment in community educations to gain support for such
127 programs.

128 The desire among urban communities and government authorities for more ecological sustainable pest
129 management is driving the need for new regulatory approaches and a shift towards IPM which uses
130 complementary techniques to limit pest levels below a given threshold, rather than focusing on
131 individual pest species (Coll and Wajnberg 2017). There is huge potential for the use of IPM to
132 control arthropod pests in urban ecosystems (Olkowski, Olkowski et al. 1976, Sawyer and Casagrande
133 1983, Brodeur, Abram et al. 2017) and in the last 30 years there has been some evidence of success
134 (Olkowski, Olkowski et al. 1978, Brewer and Stevens 1983, Wan, Cai et al. 2018) but limited

135 translation into practice. The lack of widespread adoption is likely the result of a broad range of
136 knowledge, social and environmental barriers for multiple stakeholders including tenants, landholders,
137 community groups, local government, and businesses.

138 The primary benefit of using IPM approaches to manage arthropod pests in urban areas such as
139 houses, businesses and green spaces is to reduce dependence on chemical control methods and reduce
140 the subsequent threats to the environment and human health. Research in the 1980s showed that the
141 amount of pesticides applied per area in urban gardens and lawns was higher than that used in
142 agriculture (Dahlsten and Hall 1999). Yet the current extent of this phenomenon is unknown, because
143 many countries do not keep pesticide use records (Voldner & Li 1995); in those that do, urban
144 homeowners are rarely required to have a chemical application license, or to keep records of amounts
145 used (e.g. Bekarian et al. 2006; The State of Victoria 2017). Many insecticides are broad spectrum,
146 meaning that non-target arthropods, including beneficial species, are adversely affected. This can have
147 significant impacts on ecosystem services and function (Chagnon, Kreutzweiser et al. 2015). The
148 repeated use of the same insecticides has the additional risk of increasing resistance by pests to those
149 insecticides (Naqqash, Gökçe et al. 2016). Resistance means that pest resurgences can become more
150 common and more difficult to manage, resulting in increases in the application rate and concentration
151 of insecticides in an effort to control the pests. The development of resistance results in the insecticide
152 no longer able to control the pest, posing significant challenges where there is a reliance on
153 insecticides to stop outbreaks of vector-borne disease (Deming, Manrique-Saide et al. 2016). Targeted
154 and judicious use of insecticides is essential for resistance management, which lowers the rate of
155 insecticide resistance and increases the sustainability of management of public health threats
156 (Endersby-Harshman, Wuliandari et al. 2017).

157 IPM can be more effective at controlling pests than non-integrated strategies, such as managing pests
158 in isolation or by using only broad-spectrum pesticides (Brenner, Markowitz et al. 2003, Miller and
159 Meek 2004) especially in high-density housing developments. Better land management can increase
160 urban predators and decrease reliance on chemical control in urban agriculture (Wan, Cai et al. 2018).
161 IPM strategies also have a number of flow-on benefits. For example, nursery sites for biological

162 control agents (e.g. insect predators) such as conserved vegetation and habitat restoration, can benefit
163 a variety of other species (Begg, Cook et al. 2017). In addition, highlighting the role that healthy
164 ecosystems play in IPM could increase the social and economic value of urban wildlife and help to
165 drive natural resource management policy. By educating members of local communities on pest
166 management through biological control and encouraging them to adopt IPM strategies, there is the
167 added benefit of increasing engagement with nature thereby positively influencing health (Bowler,
168 Buyung-Ali et al. 2010) and creating a drive to develop sustainable cities. Ultimately, urban IPM
169 requires a multidisciplinary approach with contributions from ecologists, social scientists, agricultural
170 scientists, ecotoxicologists, public health officials, local government, strata management, natural
171 resource managers, pest control operators and residents. Managing the expectations of so many
172 stakeholders is one of the primary barriers to IPM in cities but the ecological and social benefits are
173 invaluable.

174 3. Definition and types of IPM

175 IPM was originally developed because excessive use of broad spectrum insecticides had triggered
176 resistance among pests (rendering the insecticides ineffective), environmental damage and public
177 health concerns. In agriculture, IPM integrates a range of biological, cultural and chemical techniques
178 to maintain pest numbers below their economic threshold (Stern et al. 1959; Kogan 1998, Wilson et al
179 2018). The economic threshold of a pest is that at which the pest starts to decrease the profitability of
180 the crop. In this paper an urban pest is defined as having a monetary cost in terms of public health,
181 spoiled food, or damaged property. Therefore, if the pest in an urban setting is below its economic
182 threshold, no control is needed. The aim of IPM is not to eradicate pests, but to manage them.

183 Effective IPM incorporates three general aspects: management practices, monitoring, and decision-
184 making (Kogan 1998; Barzman et al. 2015; Wilson, Whitehouse et al. 2018). 1) Management
185 practices are in place to reduce the chance of pests exceeding their economic threshold and their need
186 to be controlled; 2) Pest numbers are monitored to identify when they reach threshold; and 3) If the
187 threshold is reached, decisions are made to select the control method which has the lowest toxicity
188 and fewer adverse impacts. Preference is given to control methods that are highly specific to the pest

189 targeted. If a pest is still above its threshold after the application of a control method, an alternative
190 method may be selected by considering its potential impact on the targeted pest and the risk of
191 adverse disruption to the ecosystem. Broad spectrum insecticides are more likely to have an adverse
192 impact on the local ecosystem because of their effects on non-targeted beneficial or neutral organisms.
193 Consequently, within IPM, broad-spectrum insecticides should be considered only when no other
194 methods have worked.

195 Techniques used in IPM exist on a continuum from prevention using management techniques (those
196 in place to stop pests reaching their economic threshold) through to intervention using pesticides
197 (those used to control pests once they do reach the threshold; Figure 1.) Control methods to stop pests
198 reaching an economic threshold include habitat manipulation to stop pests thriving (such as removing
199 stagnant water to control mosquitoes (Dale, Knight et al. 2014)); and to increase the abundance of
200 beneficial species (e.g. parasitoids, predators, pest-specific diseases or fungi, insectivorous birds). Pre-
201 emptive control methods may include introducing exotic beneficial species to a region to control pest
202 numbers (after appropriate assessment of the exotic species' potential non-target impacts (Davies,
203 Carr et al. 2011). An effective IPM system is one where pests are well monitored, and rarely reach
204 their threshold, with minimal disruption to ecosystem function.

205 If pests exceed their threshold, the methods used for control range from biological, biopesticides,
206 through to selective and broad-spectrum chemical insecticides. The softest option is augmented
207 biological control, where a key beneficial organism is released to attack pests (van Lenteren 2012).
208 This approach tends to work best in stable environments such as greenhouses, or perennial crops
209 (Stern et al. 1959; Chambers, Wright et al. 1993, Gerson and Weintraub 2012). Among the
210 biopesticides, semiochemicals alter pest behaviour rendering them less destructive (Dicke and Sabelis
211 1988, Cook, Khan et al. 2007) and microbial biopesticides contain a virus or bacterium, which attack
212 the target pest, e.g. Nuclear Polyhedrosis Virus used to control caterpillars (Washburn, Trudeau et al.
213 2003). Chemical insecticides range from those highly selective and targeted with low toxicity to
214 vertebrates, to broad spectrum chemicals, such as organochlorines and organophosphates, which are

215 highly toxic to most arthropods and vertebrates. An example of a targeted insecticide is pyriproxyfen,
216 a growth regulator that targets insect larvae by stopping them from developing into a mature insect.
217 Although good selective tools are vital in an IPM program, equally important are the pest monitoring
218 techniques. To be effective, an IPM system needs monitoring methods that are easy to undertake,
219 coupled with clear economic thresholds and a clear understanding of what can influence these
220 thresholds. Additionally, a support information network for the pest controllers is vital, so that the
221 controller can have confidence in the decisions they make. In summary, the success of an IPM
222 program depends on the pest control tools available, the dedication of the pest controllers, the
223 accuracy of the monitoring methods, and the ability of social networks to provide unbiased support to
224 the pest controllers.

225 4. Comparison of pest management in urban and agricultural systems

226 In urban environments the primary land uses that require arthropod management are houses and
227 businesses, green spaces such as parks and gardens (including water bodies), and urban agriculture
228 spaces and nurseries. There are fundamental biotic, abiotic, and social differences between these
229 urban land uses and agricultural systems that affect arthropod pests and their management. These
230 differences must be recognised in order to successfully apply IPM strategies in urban areas.

231 4.1 Biotic

232 Most agricultural IPM strategies focus on monoculture systems, where one or a few plant species are
233 grown in high density. In contrast, urban environments tend to have higher plant species richness
234 (Wania, Kühn et al. 2006, McKinney 2008), including a large proportion of exotic species. Since
235 much of urban land is privately owned and managed, socioeconomic variables can strongly influence
236 biotic diversity. For example, a study in Phoenix (USA) found higher plant diversity in wealthier
237 regions (Hope, Gries et al. 2003), while a study in Chicago (USA) found that cultural diversity
238 positively correlated with plant species richness. That is, ethnically mixed neighbourhoods had greater
239 species richness than ethnically homogeneous neighbourhoods (Lowenstein and Minor 2016).

240 Urban areas are characterised by extreme levels of habitat heterogeneity as many different land uses
241 and planting strategies can occur over relatively short spatial scales. As a consequence, species
242 turnover (beta diversity) tends to be relatively high in urban areas, with big variation in species
243 composition between adjacent but differentially managed properties (Rebele 1994, Niemelä 1999,
244 Wania, Kühn et al. 2006, Aronson, Handel et al. 2015). The high species turnover found in urban
245 areas could pose unique challenges for urban IPM because pest species are less concentrated thus
246 hindering attempts to use low-dispersal biological control agents such as mites or small parasitoid
247 species. High heterogeneity and a big proportion of privately owned (and potentially inaccessible)
248 land also makes monitoring more difficult than in traditional agricultural environments. However,
249 studies have demonstrated that where urban agriculture is undertaken in cities, it contributes to
250 increased biodiversity that may provide crucial ecosystem services such as pollination and pest
251 control (Lin, Philpott et al. 2015). Rather than a challenge, the diversity of arthropod habitats in cities
252 may offer benefits not yet fully appreciated.

253 The composition of arthropod communities differs significantly between urban and agricultural
254 landscapes (McIntyre, Rango et al. 2001). In Arizona (USA) arthropod predators, herbivores and
255 detritivores were more abundant in agricultural sites than in urban ones, but omnivores enjoyed equal
256 abundance across all site types (McIntyre, Rango et al. 2001). Urbanisation affects trophic dynamics
257 by shifting control of food web dynamics. A study in the Phoenix area (USA) found that control of
258 arthropods shifted from bottom up, resource based control in natural environments to top-down
259 control from birds in urban areas (Faeth, Warren et al. 2005). Human activities result in a large influx
260 of food waste including resources such as high sugar and high protein foods that are naturally difficult
261 to obtain. The large amounts of food waste generated by urban environments is a boon to omnivores,
262 particularly those with a large diet breadth. For example, Penick, Savage et al. (2015) found that a
263 significant proportion of urban ants' diets were obtained from 'human foods'.

264 4.2 Abiotic

265 The modification of soil structure and vegetation composition in agricultural systems results in
266 significant changes to the abiotic environment. These changes are even more extreme in urban

267 landscapes that are characterised by large areas of impervious surfaces such as bricks, pavement and
268 asphalt, in addition to the removal of shrubs and dead wood (Marzluff and Ewing 2001). Up to 80%
269 of land in central urban areas is covered by pavement or brick (Blair and Launer 1997). The high
270 concentration of impervious surfaces can lead to the ‘heat island effect’, which can cause urban areas
271 to be as much as 12 degrees C higher than surrounding areas (Imhoff, Zhang et al. 2010). These
272 abiotic changes are likely to impact pest communities in cities. In Raleigh, North Carolina (USA),
273 higher urban temperatures were correlated with increased abundance of herbivorous arthropods, but
274 not with abundance of spiders; indeed, spider taxa known to control herbivorous arthropods were
275 absent from the hottest sites (Meineke, Holmquist et al. 2017). This decoupling of predator-prey
276 dynamics can cause a loss of top-down control. Heavy metal contamination in urban soils can also
277 have a negative impact on natural enemies resulting in shorter life spans, slowed development and
278 reduced reproduction (Gardiner and Harwood 2017).

279 Urban waterways also have the potential to represent important pest insect populations because
280 mosquito species that are highly tolerant of polluted waterbodies are able to thrive (Kassim, Webb et
281 al. 2013). Without the presence of local aquatic predators or competitors more suited to less polluted
282 habitats, urban wetlands can be a source of mosquitoes (and other flying insects) that may be nuisance
283 pests or public health risks (Russell 1999).

284 4.3 Social

285 In agricultural systems, the stakeholders are primarily producers, who have similar goals in regards to
286 pest control and land management. In comparison, the complexity and diversity of stakeholders in
287 cities can make urban IPM particularly challenging. Cities contain a multitude of stakeholders often
288 with radically divergent views on how urban animals should be managed. For example, urban
289 beekeeping (of European honey bees, *Apis mellifera* L.) is being increasingly adopted by individuals
290 wanting to contribute to bee conservation efforts; yet the species is introduced in many parts of the
291 world and increasing urban beekeeping efforts may have numerous social and ecological impacts
292 (Colla and MacIvor 2016). An organic community garden, a preschool, a restaurant and a private
293 garden may have entirely different philosophies about pest management. The threshold level before
294 chemical control begins will also differ substantially between individuals, and convincing

295 stakeholders to accept non-zero thresholds may be difficult. For example, Vetter and Hedges (2018)
296 point out that homeowners would likely call a pest controller if a single venomous brown recluse
297 spider was observed in their home, despite the fact that the threshold for the safe number of brown
298 recluses is set at 6 spiders per house (Vetter and Hedges 2018). Since much of urban land is privately
299 owned and managed, landscape level IPM programs require the cooperation and coordination of many
300 independent home and business owners.

301

302 5. Examples of successful strategies using IPM to control arthropod pests

303 IPM in urban settings involves broad and complex issues and multiple types of pests, for example
304 horticultural, medical and household (Raupp et al. 2010; Fonseca et al. 2013). To understand some of
305 these complex issues it can be useful to compare examples of successful implementations in different
306 contexts.. The two case studies below of IPM in an agricultural (Australian cotton production
307 systems) and an urban setting (mosquito control) demonstrate how different environmental and social
308 challenges can be overcome in order to implement programs that are successful over long time
309 periods. We include an agricultural example because most knowledge of IPM success comes from
310 agricultural systems and this knowledge is essential to inform the design and adoption of urban IPM
311 programs. These two examples also highlight the breadth of social, economic and ecological issues
312 involved in IPM problems.

313 5.1 Pest control in Australian cotton

314 In 1999 the Australian cotton industry changed from trying to eradicate pests, to adopting an IPM
315 approach. The change was triggered by resistance by key pests to broad spectrum insecticides. In the
316 heavy pest season of 1998/99, broad spectrum insecticides were unable to control these pests,
317 rendering cotton production unprofitable, nearly collapsing the industry and triggering change.
318 Consequently, in 1999, it was the growers themselves instigating the move towards IPM (Wilson,
319 Whitehouse et al. 2018).

320 Aiding the uptake of IPM was the extensive IPM toolkit that had been developed over the decades
321 prior to 1999, during periods when resistance rendered insecticides ineffective (Wilson, Whitehouse

322 et al. 2018). The tools included 1) cultural techniques, such as plant compensation (Hearn and Room
323 1979), 2) enhancing the activity of “beneficials” (e.g. predators, parasitoids, or diseases, Mensah
324 2002) and identifying the economic thresholds for key pests (Whitehouse 2011); 3) mechanical
325 techniques including cultivating to destroy pests (Fitt and Forrester 1987); 4) targeted insecticides
326 such as spinosad (Wilson, Whitehouse et al. 2018)) and 5) effective biopesticides (Gould, Anderson et
327 al. 1991). Transgenic “Bt” cotton, which in its current form is cotton that has been genetically
328 modified to produce three proteins toxic to the key lepidopteran pests *Helicoverpa* spp, was first
329 introduced in 1996. Bt cotton, which is supported by a comprehensive Resistance Management Plan,
330 has been a corner stone of IPM in cotton since 1999 (Torres, Ruberson et al. 2010, Whitehouse,
331 Wilson et al. 2014). Since 1999, insecticide use in cotton has dropped from about 30 insecticide
332 sprays per season, to around 3 (Wilson et al 2018) illustrating the positive effect of cotton IPM.

333 Another important factor in the success of IPM in cotton has been the large support network providing
334 information to growers (the controllers in an IPM program, Figure.2). The network includes industry
335 extension officers, which distribute information from the researchers to the growers and consultants.
336 Growers are also supported by consultants (funded directly by the grower) who provide information
337 on pest management and agronomy. This network was developed through an industry wide extension
338 program which included field days, workshops, and the development of networks to enable growers to
339 discuss IPM issues and co-ordinate pest management activities across a region. Researchers working
340 with growers also demonstrated the profitability of an IPM approach (e.g. (Whitehouse 2011))
341 providing growers with more confidence to use IPM methods.

342 Key to the success of IPM in cotton, and in other agricultural systems (Bottrell and Schoenly 2018), is
343 that it was grower driven, there was an extensive and varied IPM toolkit that was effective, and a
344 large and flexible support network that not only provided information and advice to growers but
345 support within the grower community.

346 5.2 Mosquito control in urban areas

347 Mosquito-borne disease has been a growing concern for health authorities across Australia for many
348 decades. In recent years, the pest and public health risks associated with mosquitoes in urban areas

349 have increased as new residential and recreational areas encroach on existing mosquito habitats and
350 the construction and rehabilitation of wetlands enhanced conditions for mosquitoes within
351 metropolitan regions (Claflin and Webb 2017, Crocker, Maute et al. 2017). This increased exposure
352 has brought with it an increased risk of endemic mosquito-borne diseases, including the most common
353 mosquito-borne disease in Australia, Ross River virus (approximately 5,000 cases reported annually;
354 (Claflin and Webb 2015)). In addition, there is the threat of container-inhabiting mosquitoes (i.e.
355 mosquitoes adapted to exotic water-holding containers around houses) associated with the
356 transmission of dengue viruses (DENV). The container-inhabiting mosquitoes have required local
357 authorities to broaden their focus from local wetland habitats to suburban backyards (Webb and
358 Doggett 2016).

359 Mosquito management by authorities in urban settings has followed an IPM approach, both
360 historically and contemporarily. Key to success has been a diverse IPM toolkit including the use of
361 mosquito control agents, habitat management, community education and consideration of emerging
362 technologies. For example, historically, mosquito control in urban areas of Australia used a range of
363 techniques, focusing on modifying habitats, applying control agents (e.g. petroleum oils,
364 organochlorines, organophosphates) or releasing invasive fish (e.g. the plague minnow, *Gambusia*
365 *holbrooki*) (Webb and Joss 1997, Pyke 2008). However, although these techniques were diverse, they
366 were not very selective.

367 In recent decades, a more ecologically sustainable approach to mosquito control is in place,
368 underpinned by monitoring programs of mosquitoes and research identifying mosquito habitats, the
369 environmental conditions that drive mosquito abundance, and the activity of mosquito-borne
370 pathogens. Modern mosquito control primarily involves the use of more selective mosquito control
371 agents (i.e. minimal non-target and negligible human health impacts) than the previously employed
372 insecticides (e.g. organochlorides, organophosphates), including insect growth regulators (*s*-
373 methoprene); biopesticides (*Bacillus thuringiensis israelensis* (Bti)) and is complemented by
374 appropriate habitat modification (Russell and Kay 2008).

375 Mosquito IPM is constantly evolving from a social perspective. First, there are two dominant types of
376 mosquitoes: wetland mosquitoes and container-inhabiting mosquitoes. Management of these two
377 types needs to occur at two different community levels: either by local authorities (in the case of
378 wetland mosquitoes) or by householders (container-inhabiting mosquitoes). Therefore, successful
379 mosquito IPM of these mosquitoes will look quite different. In addition, concern about mosquitoes is
380 also divided between the local authorities and householder levels. Household impetus to control
381 mosquitoes will vary between regions depending on their perceived risk, irrespective of whether the
382 mosquitoes originate from wetlands or containers. Therefore, the community involved in managing
383 the mosquitoes (either the authorities or household) may be different from the community concerned
384 about their control (which again could be either the authorities or household).

385 IPM can be used both by authorities to manage wetland mosquitoes, and by households to reduce
386 container-inhabiting mosquitoes. Local authorities must co-ordinate management approaches and
387 acknowledge that mosquitoes and their habitats do not adhere to government boundaries. This
388 approach enables local governments to share resources in managing local mosquito risks, including
389 public health messaging that educate the community on personal protection measures (e.g.
390 behavioural change; the safe and effective use insect repellents; (Webb 2015, Webb and Hess 2016).
391 Better multiagency coordination will further remove barriers to effective mosquito management
392 (Dale, Knight et al. 2014).

393 Management of container-inhabiting mosquitoes is essential in order to reduce local outbreaks of
394 disease for example DENV, which is triggered when infected travellers are bitten by local *Aedes*
395 *aegypti*. This mosquito is closely associated with water-holding habitats found in backyards. In central
396 and far north Queensland, Australia, a program integrating surveillance of mosquito populations and
397 human disease, targeted insecticide use, education of the community, and employment of emerging
398 technologies, has greatly reduced the burden of disease within the region. As pest management had to
399 occur at the household level, local authorities educated householders (increasing their perceived risk)
400 by continually running awareness and education programs. These encouraged householders to
401 minimise potential mosquito habitats within their properties and enact targeted mosquito control

402 around properties of infected individuals. As householders play an important role in reducing the
403 availability and suitability of water-holding containers for *Ae. aegypti* (Kolopack, Parsons et al. 2015),
404 the strong community engagement and education strategies have been critical to the success of the
405 mosquito control program in this region.

406 5.3 Comparison of successful IPM case studies

407 Despite their differences, the examples of successful IPM in cotton pest management and mosquito
408 control reveal strong similarities. In both situations, there is an extensive toolkit ranging from habitat
409 modification, to targeted or broad-spectrum chemicals. In addition, even though tolerance thresholds
410 of pests were quite different, the use of control methods in both situations were based on monitoring
411 for pest or disease outbreaks. Key in both systems is the social aspect. First, the implementers of IPM
412 were mobilized either through a crisis (cotton growers) or through education (householders). Second,
413 there was a strong network of services supporting the controllers (Figure 2), in the form of industry
414 groups (cotton) or local authorities (mosquito management). Consequently, IPM can be effective in
415 most situations, provided controllers are motivated, a large range of management techniques are
416 available, good monitoring methods are established, and there is strong community support.

417 6. IPM strategies for different interest groups in urban areas

418 There are a number of IPM tools that have the potential to be successful in urban areas (Figure 1) but
419 the large and diverse stakeholder contingent is one of the most challenging aspects of implementing
420 IPM in cities. Urban stakeholders have different goals, values, resources and concerns; to be effective,
421 an IPM strategy must therefore be tailored toward the target stakeholder group (Schoelitz, Poortvliet
422 et al. 2018). For example, cockroaches are a nuisance pest for both home and restaurant owners, but
423 for restaurants, the presence of even low cockroach numbers can violate health regulations and incur
424 associated fines. In contrast, the economic consequences of a low-level cockroach infestation in a
425 private dwelling are substantially lower. Thus, while both restaurants and home owners would do well
426 to focus on behavioural changes to avoid cockroaches (such as ensuring all food is sealed, emptying
427 bins before night fall and locating and sealing potential cockroach harbourages) home owners, to
428 whom the economic impact is less, can wait longer before escalating to chemical control.

429 Alternatively, restaurant owners may need to implement chemical control alongside behavioural
430 changes.

431 Different stakeholders operate at very different scales and in different locations. Homeowners, for
432 example, are generally working at the relatively small scale of a privately owned dwelling while a city
433 council will need to work at the landscape scale. IPM strategies also need to be adapted to either the
434 indoor or outdoor environment. Biological control agents will be less desirable for stakeholders
435 treating indoor spaces, while insecticidal baits are best applied indoors where they are less likely to
436 have non-target impacts. Indoor pest management entails treatment in an enclosed space and thus
437 caution must be used when selecting aerosol insecticides. IPM approaches to indoor pest arthropods,
438 such as cockroaches, can reduce the use of pyrethroid insecticides while still achieving satisfactory
439 reductions in pest arthropod abundance (Zha, Wang et al. 2018).

440 While different stakeholder groups require different IPM solutions (Figure 1b), there can be
441 tremendous within-group variation. Hope, Gries et al. (2003) found that wealthier city-dwellers lived
442 in neighbourhoods with higher plant diversity; this in turn could mean a more robust guild of predator
443 and parasitoids and better natural pest control. In the United States, people living in poor
444 neighbourhoods are more likely to report the presence of cockroaches, yet this is the demographic
445 with the least access to quality pest control services and the highest rates of cockroach-sensitive
446 allergic asthma (Olmedo, Goldstein et al. 2011, Camacho-Rivera, Kawachi et al. 2014, Do, Zhao et al.
447 2016).

448 Beyond pest control within residential dwellings, there may be distinctly different attitudes to reduced
449 insecticide use and adoption of an IPM approach, depending on an individual's or community group's
450 attitudes. For example, those with an interest in improving urban biodiversity conservation (Dearborn
451 and Kark 2010) may have different motivations that lead to variability in their IPM approach. A
452 growing interest in IPM within a community may be fostered through strategies to conserve arthropod
453 populations, particularly pollinators (Hall, Camilo et al. 2017). Here IPM can play a critical role
454 facilitating community engagement with biodiversity conservation in cities.

455

456 7. Overcoming barriers to the adoption of IPM in urban areas

457 Many of the challenges involved in improving adoption of urban IPM are similar to those involved in
458 encouraging effective invertebrate conservation strategies (Cardoso, Erwin et al. 2011, New 2018).
459 Investment in long-term education and training strategies is essential to achieve a shift toward more
460 sustainable urban pest control strategies. This is a difficult task due to significant scientific, social and
461 regulatory obstacles that need to be overcome before effective IPM standards can be established for
462 urban pest control. In this section we outline a number of tools and initiatives that can be applied by
463 all stakeholders in order to overcome these barriers (summarised in Figure 2).

464 7.1 Lack of scientific knowledge

465 Scientific knowledge of many invertebrate fauna is severely limited, with large knowledge gaps in
466 taxonomy, ecology and life histories of most species (Cardoso, Erwin et al. 2011). This information is
467 critical to designing effective IPM strategies, especially involving natural enemies, or where
468 introduced pests establish in new regions. In addition, effective IPM should take a holistic multi-pest
469 approach, yet a lot of IPM research in agricultural systems has traditionally focused on individual
470 species (Horne, Page et al. 2008), with limited understanding of interactions between multiple
471 arthropod taxa because of their complexity (Costamagna 2009), or how pest management impacts
472 non-target species like pollinators (Biddinger and Rajotte 2015).

473 Designing evidence-based urban IPM programs requires long-term investment in a broad range of
474 research, beyond just the life histories and ecology of particular pest species. Urban systems foster
475 unique dynamics and interactions compared to agricultural or natural systems. Therefore, existing
476 knowledge of pest species in agricultural or natural systems may prove irrelevant in urban systems.
477 Yet arthropod communities in urban areas have been studied far less than birds and mammals (Magle,
478 Hunt et al. 2012). There is ample scope for more research to understand how multiple factors
479 influence pest populations and their management in urban areas. For example, the effects of urban
480 warming or climate change on pests (e.g. (Meineke, Dunn et al. 2013, Dale and Frank 2014)), how
481 municipal insecticide use influences pest populations and their natural enemies (e.g. (Szczepaniec,
482 Creary et al. 2011)), the relative importance of top-down and bottom-up effects on herbivorous pests
483 (Raupp, Shrewsbury et al. 2010), and the effect of numerous other complex factors like pollution and

484 traffic, fragmentation, mowing regimes, vacant land and more (Jones and Leather 2012, Gardiner,
485 Burkman et al. 2013). Much of the urban arthropod literature has focused on pest taxa, with relatively
486 limited knowledge of beneficial species, including predatory and parasitic arthropods and pollinating
487 species. Conservation biological control, a key tool in IPM programs, is one of the most understudied
488 ecosystem services in urban systems (Ziter 2016). Greater understanding of how urban landscape
489 management can enhance natural enemy populations is essential to develop effective IPM strategies.

490 7.2 Cultural and economic factors

491 Research on understanding the barriers to the adoption of IPM is dominated by work in agricultural
492 systems, with limited knowledge of social barriers to IPM adoption in urban areas. Effective IPM is
493 knowledge-intensive and insufficient knowledge or training, lack of technical support and costs to the
494 farmer are some of the key reasons cited for failure to adopt IPM programs on farms (Horne, Page et
495 al. 2008, Parsa, Morse et al. 2014). Factors such as cost also influence IPM adoption in urban systems
496 (Miller and Meek 2004, Paine, Millar et al. 2015); however, subjective human factors, like aesthetics,
497 health, comfort, are likely to be more important drivers of IPM adoption in urban settings compared to
498 agriculture (Sawyer and Casagrande 1983). For example, a study of professional landscapers in the
499 United States found that customer demand for ‘the perfect lawn’ was one of the primary barriers to
500 IPM adoption (Ingram, Stier et al. 2008). Media representation is also an important driver of people’s
501 attitude toward particular arthropods, even if they have limited personal experience with those taxa
502 (Sammet, Andres et al. 2015). Encouraging accurate media representation of arthropod behaviour and
503 ecology is critical to increasing public understanding, and this can be easily addressed via scientists
504 taking on public engagement and science communication roles in popular and social media (Smith
505 and Saunders 2016; Berenbaum 2017).

506 Individual attitudes to arthropods are a key underlying factor in social acceptance of urban IPM. Not
507 only will an individual’s attitudes vary, independent of actual or potential perceived pest impacts
508 (New 2015) but these attitudes are likely to vary greatly within and between communities engaged
509 with the implementation of IPM. Negative human attitudes to arthropods, particularly insects and
510 arachnids, are widespread and well-documented in the literature (Byrne, Carpenter et al. 1984,
511 Prokop, Tolarovičová et al. 2010, Last 2014) and are likely a fundamental driver of pest control

512 attitudes. For many people that fear or loathe arthropods, using pesticides regularly to kill unwanted
513 arthropods takes less effort and has immediate results compared to manual removal or investment in
514 biological control tools that may take longer to have an effect. These negative attitudes are not
515 changed quickly, and require comprehensive, long-term education and engagement programs (Last
516 2014). For example, the Master Gardener programs in North America have enabled participants to
517 develop their scientific knowledge of natural systems, including their understanding of insects and the
518 differences between pest and beneficial species (Schrock et al. 2000; Dirks and Orvis 2005).

519 In some instances, short-term responses by local authorities are required where an invasion by exotic
520 arthropods could increase the risk to public health, or cause significant damage to property or
521 vegetation. While in these instances control following IPM principles may not be desirable, or even
522 possible, to implement, the outbreaks can highlight that a benefit of long-term IPM is a reduced
523 reliance on insecticides. A study from California, USA, found that residents gave social and financial
524 support to manage arthropod pests of urban street trees using biological control rather than applying
525 insecticides (Jetter and Paine 2004). This was echoed in another study on street tree management in
526 California where the biological control of arthropod pests damaging trees held substantial economic
527 returns (Paine, Millar et al. 2015).

528 In agricultural systems, Farmer Field Schools have been successful in building ecological knowledge
529 and improving the uptake of IPM among farmers, with numerous positive social, economic and
530 environmental outcomes (Van den Berg and Jiggins 2007, Yorobe, Rejesus et al. 2011). Farmer Field
531 Schools have predominantly operated in Global South countries, providing on-the-job training for
532 farmers who use their own farms to trial the knowledge they learn with peer and expert support. A
533 similar model can be developed for urban communities. In particular, community gardens can be an
534 important tool in the long-term education and changing attitudes of urban communities. These
535 gardens provide spaces for creating social-ecological memories within the communities, and building
536 local traditional knowledge that transcend generations and personal networks (Barthel, Folke et al.
537 2010). For landscape scale effectiveness of biological control tools, collective and coordinated
538 neighbourhood action is essential to avoid the process being disrupted, e.g. by pesticide use (Paine,

539 Millar et al. 1997), and community education programs can achieve this (Olkowski, Olkowski et al.
540 1976).
541 Incentives for urban individuals to adopt IPM may also be somewhat different to incentives in
542 farming systems. For farmers, customer demand or increased market prices for environmentally
543 sustainable produce can be an important reason for adopting IPM (Mzoughi 2011). Customer/client
544 preferences may well influence the adoption of IPM in urban food industries, but individual urban
545 dwellers do not have the same incentive to reduce pesticide use to please others. There is an urgent
546 need for research into what influences understanding and adoption of IPM in urban systems, to inform
547 effective education programs. For example, Mzoughi (2011) found that ‘doing the right thing’ was an
548 even more important driver of a farmer’s decision to adopt IPM than cutting production costs, and
549 education programs that tap into human ethical or environmental concerns may prove effective at
550 enhancing IPM adoption in urban environments. As the community has increasingly adopted energy-
551 efficient and water-efficient strategies in cities (Dieu-Hang, Grafton et al. 2017), so too may come the
552 uptake of IPM to manage arthropod pests.

553 7.3 Management and community

554 Although good selective tools are vital in an IPM program, equally important are how those tools are
555 managed through the pest monitoring techniques. As the experience in agricultural settings has
556 shown, an effective IPM system needs monitoring methods that are easy to undertake, coupled with a
557 good understanding of the pest economic thresholds and what can influence these thresholds. Key to
558 the uptake of tools and their sustained use is the community in which IPM is undertaken. Controllers
559 need a trustworthy information network that quickly provides them with information on new
560 techniques and improved understanding of the pests and their thresholds, so that controllers can have
561 confidence in the decisions they make. In some urban contexts, commercial applicators may be more
562 trusted than scientific experts or government officials (Rickard 2011), so they need support and
563 incentives to act as agents for change in urban systems.

564 Unfortunately, in many agricultural systems support networks are built by chemical and seed
565 distributors which have a vested financial interest. An independent, trusted network is important for

566 an effective IPM program to develop and be sustained (Bottrell and Schoenly 2018). An IPM program
567 requires the controller to have a much deeper understanding of the system with which they are
568 working than an eradication program (Wilson, Whitehouse et al. 2018). Therefore, the controller
569 needs the community support of similar like- minded colleagues with whom they can discuss
570 challenges and remedies to give each other support in their decisions. In the cotton industry, this takes
571 the form of Area Wide Management groups (Wilson, Whitehouse et al. 2018). Providing residents
572 with this level of confidence would require a well-functioning information network and the
573 community support of other like-minded residents. Such programs also exist in urban areas to manage
574 pest arthropods. For example, in New Jersey, USA, area-wide IPM has been successful in reducing
575 nuisance-biting impacts of mosquitoes as demonstrated by social and economic benefits where such
576 programs exist (Shepard, Halasa et al. 2014).

577 Monitoring programs engaging citizen scientists, and crowd-sourced data, indicate how communities
578 could be engaged to enhance IPM in urban settings and can be a useful tool for building public
579 knowledge and behavioural change related to urban arthropod communities (e.g. (Jordan, Gray et al.
580 2011). For example, citizen science monitoring programs have been very successful in increasing
581 public understanding of urban bird ecology (Brossard, Lewenstein et al. 2005, Celia, Eleanor et al.
582 2005), and this is a promising area of research and practice for educating about IPM and urban
583 arthropod communities (e.g. (Paine, Millar et al. 1997). Such approaches can be adapted to monitor
584 arthropod pests with citizen science and crowd-sourced data already providing information on
585 invasive species in many regions (Hawthorne, Elmore et al. 2015, Maistrello, Dioli et al. 2016,
586 Palmer, Oltra et al. 2017). An additional advantage of establishing networks of “citizen scientists”
587 engaged in monitoring urban arthropod populations, is that it also provides the framework for
588 effective provision of community education (Williams, Hawthorn-Jackson et al. 2017).

589 The community driven approach has also been used with some success in agricultural systems: The
590 Queensland fruit fly (Qfly) monitoring program involved maintaining a trapping grid in orchards and
591 peri-urban areas across key horticultural regions in southern Australia that provided surveillance
592 outcomes and promoted community awareness (Sutherst, Collyer et al. 2000, Dominiak and Daniels

593 2012). There is also a precedent set by countries in the European Union such as Germany, France and
594 the Netherlands who all have laws introduced to restrict or ban the use of chemical pesticides in urban
595 green spaces (PAN 2009). IPM programs are extremely cost effective for urban landscape pest
596 control, but they require a long-term coordinated effort from government bodies, universities and the
597 public to maintain education, keep ahead of new pests, and integrate environmental and human health
598 goals (Goodell, Zalom et al. 2014). The sense of being part of a community involved in a citizen
599 science project gives people a sense of responsibility and ownership of the project. Such an approach
600 could be adopted to develop an IPM program in an urban setting.

601 [7.4 Policy and Regulation](#)

602 The decision of individuals to adopt IPM at the household or community level is influenced by public
603 policies. Indeed, public intervention to promote IPM is essential, to reduce the social and
604 environmental side effects of pesticide-intensive pest management, such as environmental
605 contamination and human health issues (Lefebvre, Langrell et al. 2015). Public intervention can also
606 promote coordinated area-wide programs to reduce reservoir populations of nuisance social insects
607 (Dimarco, Masciocchi et al. 2017). However, the success of these intervention programs also depends
608 on communication, training and education programs. For example, some states in the United States of
609 America have a law mandating IPM in schools. Yet Anderson (2015) found that not all schools
610 implement IPM, some schools still rely on reactive pest control, and in some schools there was a lack
611 of communication between facility managers and school administration. To enable IPM to be widely
612 adopted and successful, local authorities need to actively raise awareness of the environmental and
613 public health benefits of IPM and extend their involvement beyond education to economic and
614 operational support.

615 Despite a long history in research and rearing of generalist natural enemies, there is still a frustratingly
616 low uptake of augmented biological control by consumers. This is largely due to the dominance of
617 cheap pesticides as an easier option, and the lack of regulatory incentive for adopting sustainable pest
618 control (van Lenteren 2012). As discussed above, there are numerous ways that education programs
619 can increase public awareness and access to more sustainable pest control methods. But knowledge of
620 policy barriers to IPM in agriculture shows there are a variety of strategies to provide incentives for

621 IPM. These include changes to regulatory instruments that promote sustainable pest management
622 practices, taxes on pesticide use, commitment to publicly-funded research on the benefits IPM has on
623 urban communities, and improved information dissemination (Ramirez and Mumford 1995, Chandler,
624 Bailey et al. 2011, Lefebvre, Langrell et al. 2015).

625 8. Conclusions

626 Adoption of IPM in urban areas would produce a broad range of benefits, including reductions in
627 pesticide usage and conservation of biodiversity. In order to shift public and government practice we
628 need to better express these benefits. There are a range of IPM tools that could be applied to urban
629 pests, yet current levels of uptake have been limited by the complexity of the urban stakeholder
630 market. In order to improve urban pest management, the first step is to develop support and
631 information sharing networks that connect stakeholders and encourage good practice. There is a
632 critical need for more research into urban arthropod ecology and management, particularly to build a
633 better understanding of the environmental drivers of pest outbreaks and beneficial arthropod
634 communities in urban areas. Addressing the knowledge gaps and improving management requires
635 interdisciplinary support from a wide range of ecologists, biologists, economists and social scientists.
636 In order to apply this knowledge and implement sustainable practice we also need to establish better
637 connections between researchers, urban residents, natural resource managers, government bodies, and
638 pest controllers.

639

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1064 **Figure Legends**

1065 Figure 1: Pyramid of IPM techniques (the tool kit) that are used in a) agricultural systems, b)
1066 individuals and managers in urban systems. The foundation of the pyramid is low toxicity prevention
1067 techniques (green) which progress to more invasive intervention techniques (red) (modified from
1068 <https://www.epa.gov/managing-pests-schools/definition-verifiable-school-ipm>)

1069 Figure 2. Strategies and tools to aid the adoption of IPM by urban stakeholders. The stakeholders in
1070 IPM adoption in cities take on a number of roles (sometimes concurrently): Controllers, Monitors,
1071 Users, Educators, Suppliers and Regulators. While widespread adoption of IPM in cities faces a
1072 number of barriers, solutions to overcome these challenges exist (the IPM techniques), but rely on
1073 long-term coordinated collaborative efforts from research institutions, governments, businesses and
1074 residents. This figure demonstrates how IPM could adopted for each stakeholder group through the
1075 use of support (green), toolkits (red) and monitoring (yellow) techniques.